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Lianas Decelerate Tropical Forest Thinning During Succession

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# Abstract

The well-established pattern of forest thinning during succession predicts an increase in mean tree biomass with decreasing tree density. The forest thinning pattern is commonly assumed to be driven solely by tree-tree competition. The presence of non-tree competitors could alter thinning trajectories, thus altering the rate of forest succession and carbon uptake. We used a large-scale liana removal experiment over 7 years in a 60- to 70-year-old Panamanian forest to test the hypothesis that lianas reduce the rate of forest thinning during succession. We found that lianas slowed forest thinning by reducing tree growth, not by altering tree recruitment or mortality. Without lianas, trees grew and presumably competed more, ultimately reducing tree density while increasing mean tree biomass. Our findings challenge the assumption that forest thinning is driven solely by tree-tree interactions; instead, they demonstrate that competition from other growth forms, such as lianas, slow forest thinning and ultimately delay forest succession.

# INTRODUCTION

The successional development of ecosystems following disturbance is a foundational process in ecology (Clements, **1916**; Gleason, **1917**). Following disturbance, there is an initial pulse of tree recruitment, after which landscape-level biomass increases rapidly as trees grow in the high-resource environment (Brown & Lugo, **1990**; Finegan, **1996**). As trees grow, however, competition for increasingly scarce resources (e.g. light, water, nutrients) results in greater tree mortality and a decrease in tree density, ultimately resulting in a forest dominated by large survivors (Finegan, **1996**; Peet & Christensen, **1987**; Rees et al., **2001**; Rozendaal & Chazdon, **2015**). The inverse relationship between mean tree size (mass) and density is commonly known as self-thinning (Pickard, **1984**; Reineke, **1933**; Westoby, **1984**; White, **1981**; Yoda, **1963**).

The process of self-thinning dependably describes forest stand dynamics during succession; it has been used to identify forest biomass accumulation with the decrease in tree density and it is generally accepted as an intrinsic process to many managed and unmanaged plant communities (van Breugel et al., **2006**; Chazdon et al., **2007**; Estrada-Villegas, Bailón, et al., **2020**; Guo & Rundel, **1998**; Liu et al., **2019**; Sea & Hanan, **2012**). The concept of self-thinning was originally applied to intraspecific competition among similar-sized tree cohorts during the development of monodominant temperate forest stands (Harper, **1967**; Pretzsch, **2006**; Puntieri, **1993**; Reineke, **1933**; Westoby, **1984**; Yoda, **1963**). When the (log) mean tree biomass is plotted against the (log) tree density, there is an upper boundary for populations undergoing density-dependent mortality that represents the maximum achievable packing of trees on a surface at specific densities (Reineke, **1933**; Yoda, **1963**). This boundary has been traditionally characterised by a power-law with an exponent of −3/2 (Yoda, **1963**). More recent studies on ecological scaling, focusing on how individuals use resources as a function of their size and including a mix of tree sizes, ages and species, suggest that a slope of −4/3 is more appropriate than a slope of −3/2, which is based purely on geometrical considerations (Enquist et al., **1998**; Enquist & Niklas, **2002**). The process of self-thinning has been largely criticised when applied to natural communities (see Midgley, **2001**; Reynolds & Ford, **2005**; Weller, **1989**), where interspecific competition, or ‘alien thinning’, also takes place (Harper, **1967**). Nevertheless, whether the slope of the thinning line is close to −3/2 or −4/3, the thinning process affects most plant communities (Bazzaz & Harper, **1976**; Cousens & Hutchings, **1983**; Harper, **1967**; Rivoire & Le Moguedec, **2012**; Sea & Hanan, **2012**). Based on the scope of our study, we hereinafter refer to the natural (i.e. unmanaged) process of self-thinning (or alien thinning) in natural communities as forest thinning.

Most studies of thinning in natural forests have focused on competition among trees as the main cause of the thinning pattern, ignoring other strong competitors. However, any factor that reduces tree growth and survival (i.e. competition, stress or disturbance (Grime, **1977**)) could alter tree forest-thinning trajectories, potentially reducing the slope of the thinning relationship and thus the rate at which forests mature and sequester carbon (Deng et al., **2006**; Morris, **2003**; Zhang et al., **2011**).

In tropical forests, lianas dramatically reduce tree growth and biomass accumulation (van der Heijden et al., **2013**) and thus may alter the thinning slope attributed to tree-tree competition. Since trees comprise the majority of the biomass in tropical forest ecosystems (Gerwing & Farias, **2000**; Putz, **1983**; Schnitzer et al., **2011**), the reduction in tree growth will substantially slow landscape-level biomass accumulation (van der Heijden et al., **2013**, **2015**). Lianas are particularly abundant in secondary tropical forests (Schnitzer & Bongers, **2002**), where they tend to accumulate early in succession (Barry et al., **2015**; Dewalt et al., **2000**). While trees invest in the development of large, high-biomass trunks to support their massive crowns, lianas use those trees for support and access to the forest canopy, and thus lianas have relatively thin, low-biomass stems (van der Heijden et al., **2013**; Schnitzer et al., **2014**). Nonetheless, lianas place their leaves above those of their host canopy trees and compete intensively with trees for light (Medina-Vega et al., **2021**; Putz, **1984**; Rodríguez-Ronderos et al., **2016**), as well as for below-ground resources (Dillenburg et al., **1995**; Johnson et al., **2013**; Schnitzer, **2005**). Competition from lianas constrains tree recruitment, growth, reproduction and survival (Estrada-Villegas & Schnitzer, **2018**; García León et al., **2018**; Schnitzer & Carson, **2010**; Visser et al., **2018**); however, lianas themselves cannot sequester the quantity of carbon that they displace in their tree hosts (e.g. van der Heijden et al., **2013**; Schnitzer et al., **2014**).

Determining whether competition from lianas alters the rate of forest thinning is critical to understanding the processes governing forest recovery and succession, as well as the capacity of secondary tropical forests to uptake and store carbon. Additionally, lianas are increasing in relative abundance in tropical forests (Laurance et al., **2014**; Phillips et al., **2002**; Schnitzer et al., **2020**, **2021**), which may further alter forest thinning. Because tropical forests contain more than half of the earth's aboveground terrestrial carbon stocks (Xu et al., **2021**), and secondary forests now contribute more than one-third of all tropical forest area (e.g. Chazdon et al., **2016**), understanding the interacting factors that control the magnitude and direction of biomass accretion during secondary tropical forest succession is of critical importance to global carbon dynamics.

We used a 7-year liana removal study in a 60- to 70-year-old Panamanian secondary forest to assess whether lianas alter the thinning process in forest communities. Specifically, we tested the hypothesis that lianas reduce the slope of the thinning trajectory in tropical forests. We predicted that in the presence of lianas, the relationship between mean tree biomass and density will be shallower (i.e. less negative) compared to liana removal plots because lianas reduce tree growth rates, which reduces tree competitive effects on each other and thus limit the ability of trees to displace one another. By contrast, in the absence of lianas, trees will grow faster and thus compete more intensely, ultimately displacing each other faster, which would increase the negative slope of the thinning relationship between mean tree biomass and density (i.e. more negative). Because lianas can also influence tree recruitment and survival (e.g. Martínez-Izquierdo et al., **2016**; Perez-Salicrup, **2001**) we also quantified the separate contribution of tree recruitment and survival with and without lianas.

# METHODS

## Study sites

We conducted this study in a 60-to-70-year-old secondary forest on Gigante Peninsula, part of the Barro Colorado Natural Monument (BCNM) in central Panama. Gigante Peninsula receives ca. 2600 mm precipitation annually and has a strong dry season from January to May (Schnitzer & van der Heijden, **2019**). The forest is classified as a semi-deciduous, seasonally moist forest (Leigh, **1999**). In 2008, we established sixteen 80 × 80 m plots, and we measured the diameter, tagged, identified to species and spatially mapped all trees ≥1 cm diameter in the central 60 × 60 m of each plot. Each 60 × 60 m plot was divided into nine 20 × 20 m quadrats. In 2011, we surveyed the trees again (as well as the lianas ≥1 cm diameter) in all 16 plots and then we removed all lianas in eight randomly selected liana-removal plots, while the other eight plots were left unmanipulated as controls. We cut lianas near the forest floor with machetes and liana stems were left in the site to avoid damaging tree crowns (follows Schnitzer & Carson, **2010**). Liana-removal plots were kept free from lianas by subsequent cutting of resprouting and recruiting liana stems. We conducted additional censuses in the dry seasons of 2014 and 2018. For each census, we quantified tree growth, mortality and the recruitment of trees ≥1 cm.

## Calculation of tree biomass

We measured the stem diameter of each individual with either a caliper (stems <5 cm diameter) or a fabric diameter tape (stems ≥5 cm diameter) at 1.3 m along the stem from the rooting point on the forest floor (follows liana-specific sampling protocols by Gerwing et al. (**2006**) and Schnitzer et al. (**2008**)). Diameter measurements were collected at the beginning of the dry season for each census year and at the same point on the stem, which we marked with orange paint. We converted the DBH measurements for each tree per census year to AGB (above-ground biomass) using a regression equation derived by Chave et al. (**2014**) (See Appendix **S1**: Equation 1 in Supporting Information; see Appendix **S1**.2. for a summary of the estimated mean AGB).

## Forest thinning

We fitted thinning lines to the census data of mean tree diameter (converted to AGB kg C) and tree density per quadrat (400 m2) per survey year (2011, 2014 and 2018) and treatment (liana removal vs. control). We modeled the log-transformed (base 10) AGB as a function of the log-transformed (base 10) tree density, treatment (i.e. liana removal vs. control) and their interaction using a linear mixed-effects model assuming a Gaussian error structure (Appendix **S1**: Equation 2). We considered all living trees (including new recruits and excluding dead individuals at each surveyed year) during the seven years of manipulation, from 2011 to 2018. We also fitted independent thinning lines for each census year to assess the changes in the slopes and intercepts between treatments among censuses. There is a long-standing debate about fitting a straight line to logarithmic transformations of the original bivariate data (see Mascaro et al., **2014**; Niklas & Hammond, **2014**; Packard et al., **2011**). The debate is related to an incorrect implementation of the logarithmic transformation. Sometimes the transformation fails to linearise the observations, leading to non-log-linear allometry (Packard, **2012**), and an incorrect inference (Packard, **2014**). We checked for these potential issues with our data (Appendix **S1**.4) and also constructed a model on the original data using a lognormal error structure.

## Biomass gain from tree growth and recruitment versus biomass loss from tree mortality

The change in AGB incorporates the growth of standing trees, biomass gain from tree recruitment, and biomass loss from tree mortality. Lianas may influence tree recruitment and survival in addition to tree growth. We assessed whether a liana-induced change in the forest thinning relationship was due to differences in standing tree biomass from growth, recruitment, or mortality by fitting three independent linear mixed-effects models to the log (base 10) transformed AGB and assuming a Gaussian error structure (Appendix **S1**: Equation 3). Biomass loss from mortality can be biased towards larger stems that have the highest biomass (Nascimento et al., **2007**; Rozendaal & Chazdon, **2015**). Furthermore, lianas may negatively affect larger trees more than smaller trees because lianas tend to be in the larger trees that comprise the forest canopy (Estrada-Villegas, Hall, et al., **2020**; Lai et al., **2017**), which could lead to increasing large tree mortality. To test whether mortality varied ontogenetically between the treatments, we assessed the number of dead trees as a function of size-class and treatment by fitting a generalised linear mixed-effect model (GLMM) assuming a negative binomial error structure (Appendix **S1**: Equation 4). We defined three size classes: (1) ‘small’ (i.e. trees in the range 1 cm ≥ DBH < 5 cm), (2) ‘medium’ (i.e. trees in the range 5 cm ≥ DBH <10 cm) and (3) ‘large’ (i.e. trees with a DBH ≥ 10 cm), and included the log-transformed (base e) total number of trees per observation period as an offset (i.e. exposure variable) to adjust for the amount of opportunity for tree death.

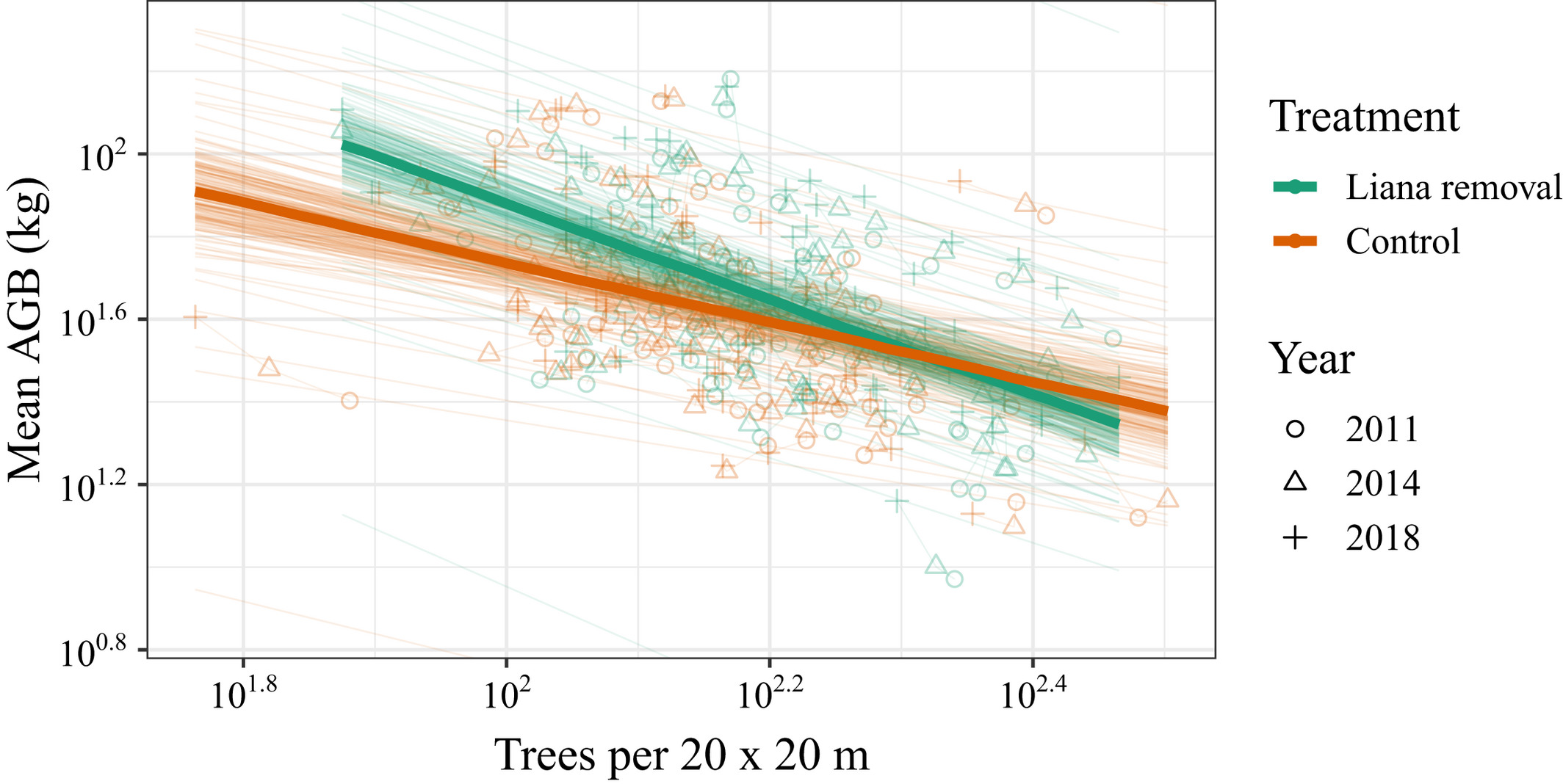
## Statistical analyses

We fitted all models in the probabilistic programming language ‘Stan’ (Carpenter et al., **2017**) via the package ‘brms’ (version 2.16.1, Bürkner, **2018**) and ‘cmdstanr’ (version 0.4.0, Gabry & Češnovar, **2021**) in ‘R’ (version 4.1.2, R Core Team, **2021**). See Appendix **S2** for the description and the sensitivity assessment of the priors used in the analyses. We estimated the coefficients of each model using four Markov chains and a number of iterations that varied per model (Appendix **S2**.2). We monitored Markov chain mixing properties and checked parameter convergence graphically via trace plots of the estimated coefficients (Appendix **S3**.1–8) and by checking the Rhat metric (Gelman et al., **2013**). The goodness-of-fit for each model was then inspected via posterior predictive model checks (Conn et al., **2018**; Gabry et al., **2019**), where simulation predictions from the best-fitted models are compared to the observed data (Appendix **S3**.9). This process allowed us to assess any obvious discrepancies between the final model and the observed data before reporting. Parameter values are presented using the median of the posterior distribution and the uncertainty in the estimates was summarised using the 95% credible intervals (CI’s) computed using the highest density interval (HDI) of posterior distributions, which favours probable over central values and is recommended for non-symmetric posterior distributions (Kruschke, **2014**).

# RESULTS

## Forest thinning

Both treatments (liana-removal and control plots) showed a strong power-law relationship between mean tree biomass and tree density (Figure **1**; Appendix **S4**: Table **S1**, model A). The negative slopes indicate that increments in mean tree AGB for both treatments (liana removal and control) were associated with reductions in the number of trees, consistent with the process of forest thinning. For the liana removal plots, the mean slope of the thinning line was −1.15 [−1.38, −0.91]. By contrast, the mean slope of the thinning relationship for the control plots, where lianas were present, was flatter (+0.44 [+0.11, +0.79]; Appendix **S4**: Table **S1**), indicating that lianas reduced the increase in mean tree biomass with forest thinning, which ultimately constrained the speed of forest-level biomass accumulation (Figure **1**). Forest thinning derived equations for liana removal and control plots are *log10 W* = *4*.*18*–*1*.*15 log10 N* and *log10 W* = *3*.*14–0*.*71 log10 N*, respectively, where *W* is the mean weight of trees and *N* is tree density. The model explained 96% of the variation in the data (Conditional *R*2 = 0.96 and Marginal *R*2 = 0.26). A model using a lognormal error structure for the original data used to construct thinning lines showed similar results (Appendix **S4**: Table **S3** and Figure **S1**).

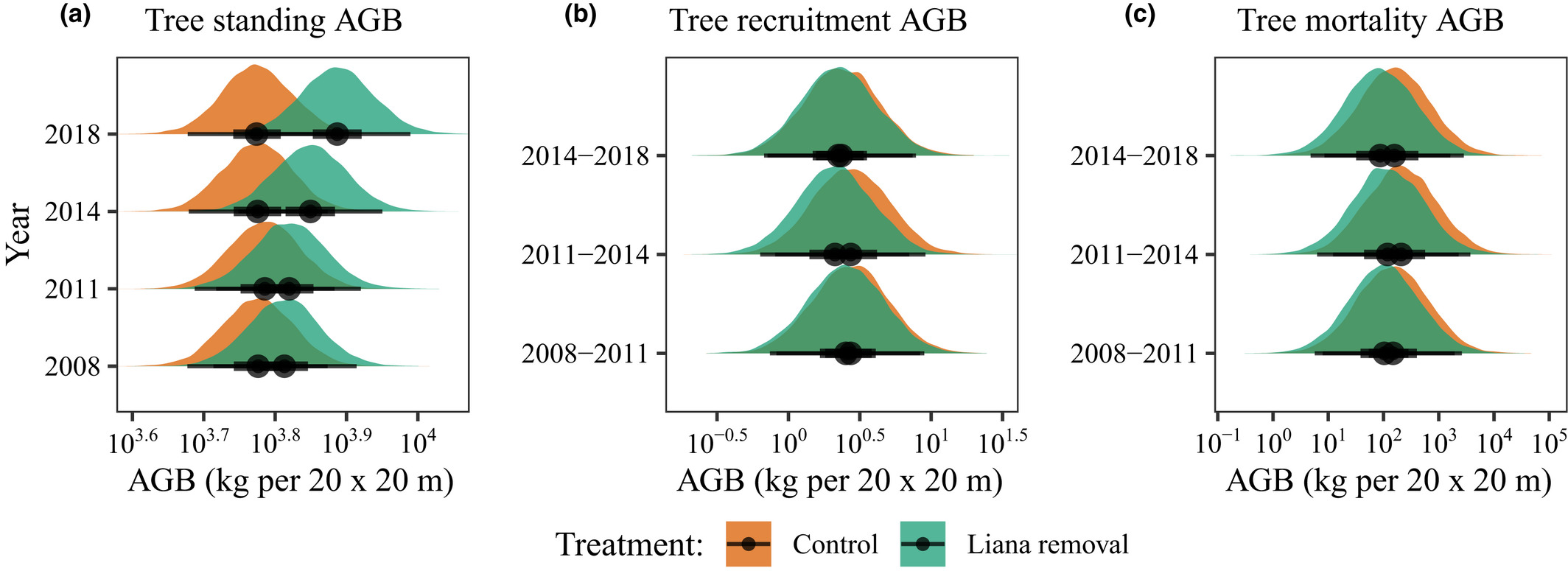


**FIGURE 1** Thinning lines for liana removal and control plots in Gigante Peninsula, Panama. The vertical axis indicates the log-transformed (base 10) AGB and the horizontal axis indicates the log-transformed (base 10) tree density (or the number of trees per 400 m2 [20 × 20 m quadrats]). Green points were used for the liana removal plots and orange for the control plots. Individual points indicate subplot or quadrat-level observations per year (shape of the point) and the lines linking the points indicate the repeated observations per subplot. The thick orange and green lines indicate the predicted median for the control and liana-removal plots, respectively, calculated from 200 draws from the posterior predictive distributions, indicated by tlight orange and light green lines that run parallel to the medians

The forest thinning relationship (i.e. the slope coefficient) was unequivocally different between treatments when all data were combined (Appendix **S4**: Table **S1**, model A and Figure **1**). Within censuses, the slope of thinning relationship did not differ between treatments, but the y-intercept was higher in the liana-removal plots in years 2014 (Appendix **S4**: Table **S1**, model C) and 2018 (Appendix **S4**: Table **S1**, model D), indicating that in the absence of lianas, mean tree AGB was increasingly greater at the same tree density than in the control plots (Appendix **S4**: Figure **S2**). There was no pre-treatment (year 2011) difference in the slope coefficient or mean tree AGB per tree density (the y-intercept) between treatments (Appendix **S4**: Table **S1**, model B and Figure **S2**a). These findings indicate that lianas reduce tree biomass accumulation for a given tree density and the effects appeared to strengthen with time.

## Standing tree biomass, biomass recruitment and biomass mortality

Lianas constrained biomass accumulation in control plots by reducing the growth of living trees, not by their effects on recruitment biomass nor mortality biomass (Figure **2**). For standing biomass, trees in control plots had lower median biomass than trees in liana-removal plots in years three (year 2014) and seven (year 2018) following the liana removal manipulation (2011) (Figure **2a**; Appendix **S4**: Table **S4**, model A). The model explained 95% of the variation in the data (Conditional *R*2 = 0.95 and Marginal *R*2 = 0.05). We did not find any differences in the gain in tree biomass from recruitment (Figure **2b**; Appendix **S4**: Table **S4**, model B), the loss in tree biomass from mortality (Figure **2c**; Appendix **S4**: Table **S4**, model C), or the number of dead trees per size-class between treatments (Appendix **S4**: Table **S4**, model D). Nevertheless, we observed higher mortality of trees in the smaller size class in both treatments (Appendix **S4**: Figure **S3**). The models for tree biomass recruitment and tree biomass mortality explained 74% (Conditional *R*2 = 0.74 and Marginal *R*2 = 0.69) and 23% (Conditional *R*2 = 0.23 and Marginal *R*2 = 0.22) of the variation in the data respectively. The model that assessed the number of dead trees per size class and treatment explained 82% (Conditional *R*2 = 0.82 and Marginal *R*2 = 0.68) of the variation in the data.



**FIGURE 2** Predicted tree (a) standing biomass, (b) biomass gain from recruitment and (c) biomass loss from mortality for control (orange) and liana removal plots (green) on Gigante Peninsula, Panama. The vertical axes indicate the census years (panel a) and inter-census periods (panels b and c). The horizontal axes indicate the predicted log-transformed (base 10) total AGB (kg) per 400 m2 (20 × 20 m). The black interval bar indicates the median (at the circle), and the 50% and 95% credible intervals calculated from 200 draws from the posterior predictive distribution

# DISCUSSION

A fundamental assumption in ecology is that the pattern of tree thinning in a forest is driven by tree-tree competition, which predicts that tree density decreases as trees increase in size during forest development (Pickard, **1984**; Reineke, **1933**; Westoby, **1984**; White, **1981**; Yoda, **1963**). That is, as trees grow, they compete intensely, ultimately displacing each other, and surviving trees increase in biomass after the loss of neighbouring trees. In the absence of lianas, we found that trees maintain high rates of thinning, demonstrating that, indeed, tree-tree competition during succession leads to a predictable loss in tree density with a commensurate increase in mean tree biomass.

However, intense competition from lianas reduced the tree thinning slope, presumably by changing competition from solely tree-vs-tree to liana-vs-tree-vs-tree. The addition of intense competition from lianas reduced tree growth (this study) and biomass uptake (van der Heijden et al., **2015**, **2019**), which slowed tree-vs-tree competition. Lianas alter the rate at which trees grow and thus displace each other, ultimately changing thinning rates in tropical forests. Our findings are consistent with previous studies that have used liana removal experiments to demonstrate that, regardless of forest type, age or geographic location, lianas have detrimental effects on tree growth (Estrada-Villegas & Schnitzer, **2018**; Marshall et al., **2017**; Mills et al., **2019**; Toledo-Aceves, **2015**; Villegas et al., **2009**), and thus we believe that the ability of lianas to slow forest thinning and thus delay forest succession is broadly applicable to tropical forests.

Our findings challenge the assumption that forest thinning is driven solely by tree-vs-tree interactions. Instead, we show that liana competition changes tree thinning trajectories. Lianas slow forest thinning by reducing tree growth, but not by altering tree recruitment or mortality; we did not find evidence for variations in forest thinning trajectories resulting from liana-induced changes in tree recruitment or mortality. The primary importance of growth to the thinning pattern is consistent with observations in Costa Rica and Mexico, where tree basal area accumulation was mostly associated with tree growth rates but not with changes in stem density resulting from recruitment and mortality (Chazdon et al., **2007**, **2010**). Therefore, the negative effect of lianas on tree growth influenced the thinning relationship between tree density and mean tree biomass.

## Liana-specific negative effects on tropical tree growth and forest thinning

The strong negative effects of lianas on tree growth and biomass increment were likely due to competition for shared resources. Both growth forms utilise the same set of resources (e.g. light, soil water and nutrients). In addition, lianas use the tree's architecture for support and access to high light positions on the forest canopy. Once in the forest canopy, lianas place their leaves over those of their host trees and access the most exposed light conditions (Avalos et al., **2007**; Medina-Vega et al., **2021**; Rodríguez-Ronderos et al., **2016**). This interaction between lianas and trees results in strong competition for light. However, lianas can also compete intensely for belowground resources (Perez-Salicrup & Barker, **2000**; Perez-Salicrup et al., **2001**; Schnitzer, **2005**; Schnitzer et al., **2005**), suggesting that there may be a similar overlap between liana and tree roots.

In liana-dominated landscapes, weaker competition between trees due to the negative effects of lianas on tree growth may result in relatively slow thinning rates and thus slower forest succession (e.g. Figure **1**). By slowing tree-vs-tree interactions, lianas may delay the displacement of early successional tree species by later-successional tree species, and thus may maintain a larger number of tree species in tropical forests. Alternatively, because the strength of the negative effect of lianas varies with tree species identity (e.g. Visser et al., **2018**), lianas may displace some species faster than others during succession, which could hasten the loss of tree species diversity during succession. Competition from other growth forms, such as shrubs or herbs, may also alter tree recruitment, but this effect appears to be temporary (Duncan & Chapman, **2003**; Frappier et al., **2004**), and it may not alter the thinning trajectory of a forest undergoing density-dependent mortality. Although lianas are a key component of tropical and temperate forests around the world, their contribution to forest dynamics, composition, and structure is most substantial in the tropics (Gentry, **1992**; Schnitzer & Bongers, **2002**, **2011**), suggesting important differences in forest succession and forest thinning between tropical and temperate forests.

## Pervasive negative effects of lianas on secondary forest carbon accumulation

By reducing the slope of forest thinning, lianas reduce the capacity for regenerating secondary forests to accumulate carbon. Our experimental findings are consistent with other studies. For secondary tropical forests in Panama, lianas reduced forest level carbon accumulation up to *c*. 22% (Estrada-Villegas, Hall, et al., **2020**; Lai et al., **2017**) and up to 76% for trees larger than 10 cm DBH (van der Heijden et al., **2015**). Lianas themselves contributed very little to the carbon they displaced (Estrada-Villegas, Hall, et al., **2020**; van der Heijden et al., **2015**; Lai et al., **2017**). The relatively small contribution of lianas to forest-level carbon results from their low stem volume (Schnitzer et al., **2012**, **2021**), slow accumulation of biomass (Letcher & Chazdon, **2009**), and their greater allocation of above-ground biomass to leaves than to the stem than similar-sized trees, which lowers their capacity to store carbon (Chave et al., **2001**; Gerwing & Farias, **2000**; Putz, **1983**).

Lianas are particularly abundant early in forest succession, and the observed negative effects of lianas on forest thinning have important ramifications for carbon uptake in regenerating tropical forests. Most regenerating tropical forests have faster growth and higher net carbon uptake than old-growth forests (Chazdon et al., **2016**; Poorter et al., **2016**). These young forests are characterised by the vigorous growth of the many light-demanding trees (Finegan, **1996**). Regenerating tropical forests also have fast biomass accumulation, high tree species diversity and high tree species composition relative to late-successional forests (Poorter et al., **2021**). Because of the high productivity of regenerating tropical forests and the increasing loss of old-growth tropical forests worldwide, secondary forests are expected to play an important role in the global carbon dynamics (Chazdon et al., **2016**; Grace et al., **2014**). However, high liana abundance in young forests (e.g. 40 years and younger) (Barry et al., **2015**; Dewalt et al., **2000**; Schnitzer et al., **2012**, **2021**) reduce tree growth and biomass accumulation (Estrada-Villegas, Hall, et al., **2020**). Therefore, the negative effects of lianas on tree-tree competition and forest thinning may be particularly important in the early stages of forest succession, where lianas likely reduce the potential of secondary forests to sequester carbon (Poorter et al., **2016**).

## Implications of increasing liana abundance for forest succession

The contribution of lianas to forest structure and dynamics appears to be increasing in tropical forests (Schnitzer & Bongers, **2011**), which may further slow tree thinning and, concomitantly, reduce forest biomass uptake. Multiple long-term studies in the neotropics and one study in South India (Pandian & Parthasarathy, **2016**) reported an increase in liana density and biomass in both absolute terms and relative to trees (Chave et al., **2008**; Ingwell et al., **2010**; Laurance et al., **2014**; Phillips et al., **2002**; Schnitzer et al., **2020**, **2021**; Wright et al., **2004**). The increase in lianas relative to trees in tropical forests suggests a greater role of these non-tree competitors in future forest succession and thinning. Moreover, among tropical forests, the negative effects of lianas on forest succession and thinning may not be homogeneous but vary with liana gradients and may become even stronger in forests where lianas are naturally more abundant or in forests that are experiencing greater increases in liana abundance (Schnitzer & Bongers, **2011**).

Among tropical forests, liana abundance and diversity peak in highly seasonal forests and decrease with increasing mean annual precipitation, increasing soil moisture availability (e.g. Manzané-Pinzón et al., **2018**), and decreasing strength of seasonal drought (DeWalt et al., **2010**; Parolari et al., **2020**; Swaine & Grace, **2007**). This unique distribution of lianas is thought to be driven by a greater ability to benefit from high dry season light availability than trees, thus resulting in higher rates of growth and survival, and ultimately greater liana abundance (Schnitzer, **2005**, **2018**). In the context of our results and the unique distribution of lianas, we hypothesise that forests with relatively strong seasonality of rainfall, where lianas are most abundant, may experience slower rates of thinning than forests with higher precipitation and lower seasonality. In wet, aseasonal tropical forests, where lianas are less abundant, forest thinning trajectories may be steeper due to less liana-tree competition and thus more intense tree-tree competition.

# CONCLUSIONS

Lianas alter forest thinning trajectories in secondary tropical forests by decreasing tree growth. The pervasive negative effects of lianas on tree growth appeared to decrease tree-tree competition and thus slowed the ability of trees to displace one another, even as individual trees grew to large sizes. Without lianas, tree growth rates were significantly higher, resulting in greater tree-tree competition and thus a greater ability of trees to suppress and displace weaker neighbours. Moreover, biomass accumulation in liana-free plots was particularly high because surviving trees more rapidly compensated for the biomass loss of dead trees. The slower accumulation of carbon that resulted from the effect of lianas on tree thinning could have substantial negative effects on tropical forest succession and the global carbon cycle, since secondary forests are a critical component of the tropical carbon budget.

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# AUTHOR CONTRIBUTION

J.A. Medina-Vega (JAM-V) and S.A. Schnitzer (SAS) conceived the conceptual idea for the manuscript. SAS led the data collection efforts. JAM-V conducted data management and the statistical analyses with comments from SAS and G. van der Heijden (GvdH). JAM-V wrote the manuscript. SAS and GvdH helped with the conceptual development and commented on the manuscript.

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